

## The potential habitats of two submerged macrophytes, *Myriophyllum spicatum* and *Hydrilla verticillata* in the river ecosystems, South Korea

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**Abstract** – We examined the environmental factors that characterize the habitats of two submerged macrophytes, *Myriophyllum spicatum* and *Hydrilla verticillata*, in South Korea using generalized additive models (GAMs). We performed a vegetation survey and measured water depth and water velocity from May to September, 2014–2015. Averaged water quality data from the Ministry of Environment’s national water quality measurement network from January 2012 to October 2015 were used for modeling. Potential habitats of *M. spicatum* were linked with chlorophyll *a*, nitrate nitrogen, suspended solids, water temperature, water depth, and water velocity (deviance explained = 28.7%, accuracy = 74%). In the case of *H. verticillata*, electrical conductivity and suspended solids were key habitat factors (deviance explained = 23.4%, accuracy = 75%). Model results were highly consistent with observations and field data. Model performances were evaluated by the accuracy rate, the area under the receiver operating characteristic curve, the kappa value, and field verification, and were in generally good agreement. Few studies have evaluated the developed models using the independent field data. By understanding the environmental factors that characterize the habitats for submerged macrophytes, our results contribute to the development of conservation and management strategies for river ecosystems.

**Keywords** : Eurasian watermilfoil / *Hydrilla* / water condition / habitat suitability / generalized additive model

**Résumé** – Les habitats potentiels de deux macrophytes submergés, *Myriophyllum spicatum* et *Hydrilla verticillata* dans les écosystèmes fluviaux, Corée du Sud. Nous avons examiné les facteurs environnementaux qui caractérisent les habitats de deux macrophytes submergés, *Myriophyllum spicatum* et *Hydrilla verticillata*, en Corée du Sud à l’aide de modèles additifs généralisés (GAMs). Nous avons effectué un relevé de la végétation et mesuré la profondeur et la vitesse de l’eau de mai à septembre 2014-2015. Les données moyennes sur la qualité de l’eau provenant du réseau national de mesure de la qualité de l’eau du ministère de l’Environnement pour la période de janvier 2012 à octobre 2015 ont été utilisées pour la modélisation. Les habitats potentiels de *M. spicatum* étaient liés à la chlorophylle *a*, au nitrate, aux solides en suspension, à la température de l’eau, à la profondeur et à la vitesse de l’eau (variance expliquée = 28,7 %, précision = 74 %). Dans le cas de *H. verticillata*, la conductivité électrique et les solides en suspension étaient des facteurs clés de l’habitat (variance expliquée = 23,4 %, précision = 75 %). Les résultats des modèles étaient très cohérents avec les observations et les données de terrain. Les performances du modèle ont été évaluées par le taux de précision, la surface sous la courbe caractéristique de fonctionnement du récepteur, la valeur kappa et la vérification sur le terrain, et elles ont été généralement bien concordantes. Peu d’études ont évalué des modèles élaborés à l’aide des données indépendantes sur le terrain. En comprenant les facteurs environnementaux qui caractérisent les habitats des macrophytes submergées, nos résultats contribuent à l’élaboration de stratégies de conservation et de gestion des écosystèmes fluviaux.

**Mots-clés** : myriophylle eurasiennne / *Hydrilla* / qualité de l’eau / convenance de l’habitat / modèle additif généralisé

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## 1 Introduction

Submerged macrophytes play an important role as a producer in food webs, shelter and forage for other organisms, and as a water quality indicator (Nieder *et al.*, 2004). In addition, submerged macrophytes produce oxygen in stagnant regions and prolong the hydrologic retention time for the removal of particulate nutrients (Nepf *et al.*, 2007). Despite the importance of submerged macrophytes, the formation of dense monotypic stands has adverse effects on the diversity and richness of invertebrates and fish (Buchan and Padilla, 2000). Dense beds of submerged macrophytes can produce organic materials from actively growing or senescing macrophytes and cause eutrophication of the water column (Chambers *et al.*, 1999). Moreover, their proliferation can impede water flow, clog inlets of reservoirs, and interfere with recreational activities (Kenneth, 1996).

The spatial distribution of organisms is related to species dispersal and survival at a regional scale. In addition, abiotic conditions (environmental constraints) and biotic interactions (*e.g.*, competition and herbivory) influence species distributions at a local scale (Austin, 2002; Bučas *et al.*, 2013; Chappuis *et al.*, 2014). The occurrence and abundance of submerged macrophytes are influenced by chemical and physical factors, such as water quality, light availability (Dennison *et al.*, 1993), water transparency, water depth (Canfield *et al.*, 1985), channel slope, channel dimensions (O'Hare *et al.*, 2011), and hydrological regime (Franklin *et al.*, 2008). Understanding how diverse environmental factors affect the habitats of submerged macrophytes is important for flow control, sediment transport (Järvelä, 2005), and assessments of the ecological condition of rivers (Clayton and Edwards, 2006).

A variety of statistical approaches such as generalized linear models and generalized additive models (GAMs) are important tools for predicting the likely occurrence or distribution of a species (Pearce and Ferrier, 2000; Austin, 2002). In particular, GAMs are used extensively in habitat suitability modeling and identification of the optimal environmental conditions for a given species; data are fitted using a semi-parametric model to predict non-linear responses to the exploratory variables (Elith *et al.*, 2006; Drexler and Ainsworth, 2013; Li and Wang, 2013). GAMs not only have a strong statistical foundation, but can be used to realistically model ecological relationships (Yee and Mitchell, 1991; Sanchez *et al.*, 2008). When the relationship between a species distribution and environmental variables is complex, GAMs are practical and perform as well or better than other types of predictive models (Drexler and Ainsworth, 2013; Li and Wang, 2013). GAMs have been used to examine potential seagrass habitats (Lathrop *et al.*, 2001; Downie *et al.*, 2013), fish production and distributions (Borchers *et al.*, 1997; Buisson and Blanc, 2008; Murase *et al.*, 2009; Solanki *et al.*, 2016), and terrestrial plant distributions (Yee and Mitchell, 1991; Austin and Meyers, 1996; Thuiller *et al.*, 2005), but few studies have used GAMs to examine submerged macrophytes in river ecosystems.

Recently, river ecosystems in South Korea have experienced channel dredging, channelization, and dam construction for flood control during rainfall periods and to secure water resources during drought periods (Woo, 2010). In particular,

the “Four Major Rivers Project” (2009–2012) involved the construction of 16 weirs and three dams in the Han, Geum, Nakdong, and Yeongsan Rivers (Lah *et al.*, 2015). This national project aimed to secure water resources, reduce flooding, improve water quality, and create multipurpose public spaces for local residents (Jun and Kim, 2011). Despite substantial controversy surrounding the effectiveness of this project (Normile, 2010), it drastically changed the natural riverine habitats and enabled the artificial manipulation of the water level and the regulation of water flow (Lee *et al.*, 2014). The modified slow velocity in regulated streams may increase the abundance of macrophytes (Bunn and Arthington, 2002) and devastate habitats for organisms adapted to the natural discharge regime (Dynesius and Nilsson, 1994).

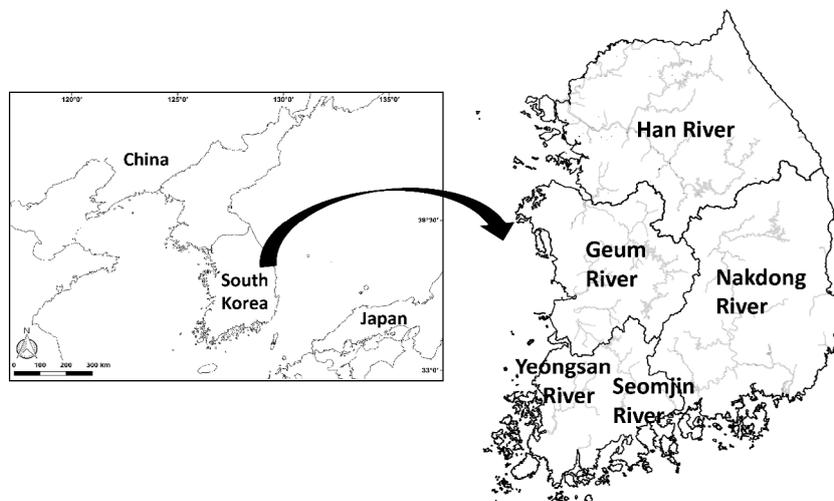
Alterations to hydrological regimes affect the structure and function of aquatic ecosystems, resulting in changes in the spatial distributions of submerged macrophytes (Tian *et al.*, 2015). Submerged macrophytes in the river ecosystems of South Korea were found to change from slow flowing streams to large rivers after weir construction. *Myriophyllum spicatum* L. (Eurasian watermilfoil) and *Hydrilla verticillata* (L. f.) Royle (Hydrilla) are native species in South Korea; however, they are fast growing and the most abundant macrophytes in Korean rivers. They are invasive species and strong competitors in Europe, the United States, and South America owing to their rapid and dense growth (Van *et al.*, 1999; Gassmann *et al.*, 2006; Beck *et al.*, 2008). These two species are cosmopolitan angiosperms with extensive worldwide ranges (Zhou *et al.*, 2016) and overrun various habitats, from lentic to lotic systems, and in turn affect flow velocity and nutrient cycling in the water column (Sousa, 2011).

Accordingly, it is necessary to understand the current distributions as well as predict suitable habitats of submerged macrophytes with high invasive potentials for river management and conservation planning. Few studies have predicted the distributions of submerged macrophytes in the rivers using GAMs (Ahmadi-Nedushan *et al.*, 2006; Camporeale and Ridolfi, 2006). Moreover, the developed GAMs have rarely been validated via independent field data (Guisan *et al.*, 2002). The aims of the study were to examine under what environmental conditions *M. spicatum* and *H. verticillata* are likely to occur in and to predict and validate their suitable habitats. Our specific objectives were (1) to characterize the relationship between environmental variables and the occurrence of two submerged macrophytes (*M. spicatum* and *H. verticillata*) using GAMs, (2) to predict the suitable habitats for these two submerged macrophytes, and (3) to evaluate GAMs by applying the independent field data.

## 2 Materials and methods

### 2.1 Study sites and data collection

The study sites were located in the central and southern parts of the Korean Peninsula (33°–39° N, 124°–130° E) which has a temperate climate. The mean annual precipitation is approximately 1300 mm, and the mean annual temperature is 10–14 °C (Korea Meteorological Administration, [http://www.kma.go.kr/weather/climate/average\\_normal.jsp](http://www.kma.go.kr/weather/climate/average_normal.jsp)). Two-thirds of the annual precipitation occurs in the summer



**Fig. 1.** Studied streams in South Korea. The streams were roughly classified into five rivers: Han River, Geum River, Nakdong River, Yeongsan River, and Seomjin River.

(June–August). The winter season (December–February) tends to be dry and cold, reflecting a typical continental weather pattern (Shin and Chung, 2011).

Surveyed streams and rivers converge into five major rivers: Han River, Geum River, Nakdong River, Yeongsan River, and Seomjin River (Fig. 1). For the vegetation survey, 197 sites in the rivers and tributaries were randomly selected, including 71 sites in the Han River, 43 sites in the Geum River, 46 sites in the Nakdong River, 27 sites in the Yeongsan River, and 10 sites in the Seomjin River. Sites were assigned to water quality monitoring towers operated by the Ministry of Environment. The presence and absence of submerged macrophytes were recorded using a belt transect (50 m × 2 m) parallel to the direction of water flow from May to September in 2014–2015 (Dawson *et al.*, 1999; Riis *et al.*, 2001) and from June to July in 2016 for field verification. The belt transect was spaced 1 m from the edge of the river. Submerged macrophytes are rarely observed at depths exceeding 2.5 m in rivers and streams (Riis *et al.*, 2001; Angradi *et al.*, 2013); examinations were performed by wading or in a boat with a rake for confirmation (Park *et al.*, 2013).

Water quality data were acquired from the Ministry of Environment's national water quality measurement network (<http://water.nier.go.kr>) from January 2012 to October 2015. Averaged water quality data calculated from monthly measurements were used for modeling. We chose 14 water environmental variables that had a postulated relationship with the ecological requirements of submerged macrophytes (Gallego *et al.*, 2015; Klippel *et al.*, 2016), namely, water temperature, pH, biochemical oxygen demand (BOD), chemical oxygen demand (COD), suspended solids, total nitrogen, ammonium nitrogen, nitrate nitrogen, total phosphorus, total organic carbon, electrical conductivity, total dissolved nitrogen, total dissolved phosphorus, and chlorophyll *a*. In each transect, water depth (meter stick) and water flow (Flowwatch; JDC Electronic SA, Yverdon-les-Bains, Switzerland) were measured three times where vegetation was most abundant, and the values were averaged. If submerged macrophytes were lacking, water depth and water flow were measured in triplicate at the midpoint of the transect. We recorded channel structure and confluence points according to *in situ* survey and stream order data obtained from the Water

Resources Management Information System (the Ministry of Land, Infrastructure and Transport, [www.wamis.go.kr](http://www.wamis.go.kr)).

## 2.2 Model building

All statistical analyses were performed in R (R Development Core Team, 2016). Prior to the statistical analysis, all explanatory variables were  $\log_{10}$ -transformed to improve normality (Chappuis *et al.*, 2014). Pearson correlation analyses were performed to detect high multicollinearity (Zhao *et al.*, 2014; Wedding and Yoklavich, 2015) and the variables that have correlation coefficient greater than 0.75 with another variable were eliminated (Kuhn and Johnson, 2013) (Tab. S1). Total nitrogen, BOD, COD, ammonium nitrogen, total phosphorus, total organic carbon, and total dissolved nitrogen were excluded because they were highly correlated with suspended solids, nitrate nitrogen, and total dissolved phosphorus. Multivariate statistical methods, such as principal component analysis (PCA), have been recommended to determine the main environmental factors before inclusion in models (Shmueli, 2010; Zhao *et al.*, 2014). Six significant elements of water environmental factors determined via PCA were selected and two elements from *in situ* measurements (water depth and water velocity) were added to the GAMs (Tab. S2). Finally, eight environmental descriptors were included in GAMs: chlorophyll *a*, electrical conductivity, nitrate nitrogen, suspended solids, total dissolved phosphorus, water temperature, water depth, and water velocity. All variables describing the physical and chemical properties of water included in the GAMs are described in Table 1.

We used GAMs to approximate the probability of taxon presence with respect to the predictors. A binomial distribution was specified (presence=1 and absence=0) with a logit link function relating the dependent variables to the predictors. Such approach allows modeling presence–absence data at survey sites, thereby enabling the probability of species occurrence to be predicted from independent data at unsurveyed sites (Pearce and Ferrier, 2000). The descriptor variables were modeled as cubic splines, with four degrees of freedom for smoothing splines (Lehmann, 1998;

**Table 1.** Mean, standard error (SE), minimum (Min), and maximum (Max) values for physical and chemical properties of water at 197 sites. Chlorophyll *a*, electrical conductivity, total dissolved phosphorus, nitrate nitrogen, and suspended solids are presented as mean values of monthly estimates from 2012 to 2015, water temperature is presented as the mean from May to October, and water depth and velocity data were obtained at the sampling date.

Variable	Unit	Mean	Median	SE	Min	Max
Chlorophyll <i>a</i>	mg/m <sup>3</sup>	14.0	9.6	0.9	0.9	65.1
Electrical conductivity	μS/cm	328	258	19	77	1515
Total dissolved phosphorus	mg/L	0.076	0.035	0.009	0.008	1.282
Nitrate nitrogen	mg/L	2.2	2.1	0.1	0.5	5.9
Suspended solids	mg/L	9.6	8.3	0.5	1.1	35.9
Water temperature	°C	23.1	23.5	0.1	17.7	26.0
Water depth	m	0.61	0.50	0.02	0.10	1.60
Water velocity	m/s	0.10	0.02	0.01	0.00	0.90

**Table 2.** Selected environmental variables and deviance explained in GAMs for *Myriophyllum spicatum* and *Hydrilla verticillata*.

Species	Environmental variable	Significance ( <i>p</i> -value)	Deviance explained
<i>Myriophyllum spicatum</i>	Chlorophyll <i>a</i>	<0.001	28.7%
	Nitrate nitrogen	0.003	
	Suspended solids	<0.001	
	Water temperature	0.034	
	Water depth	0.030	
	Water velocity	0.010	
<i>Hydrilla verticillata</i>	Electrical conductivity	<0.001	23.4%
	Suspended solids	0.010	

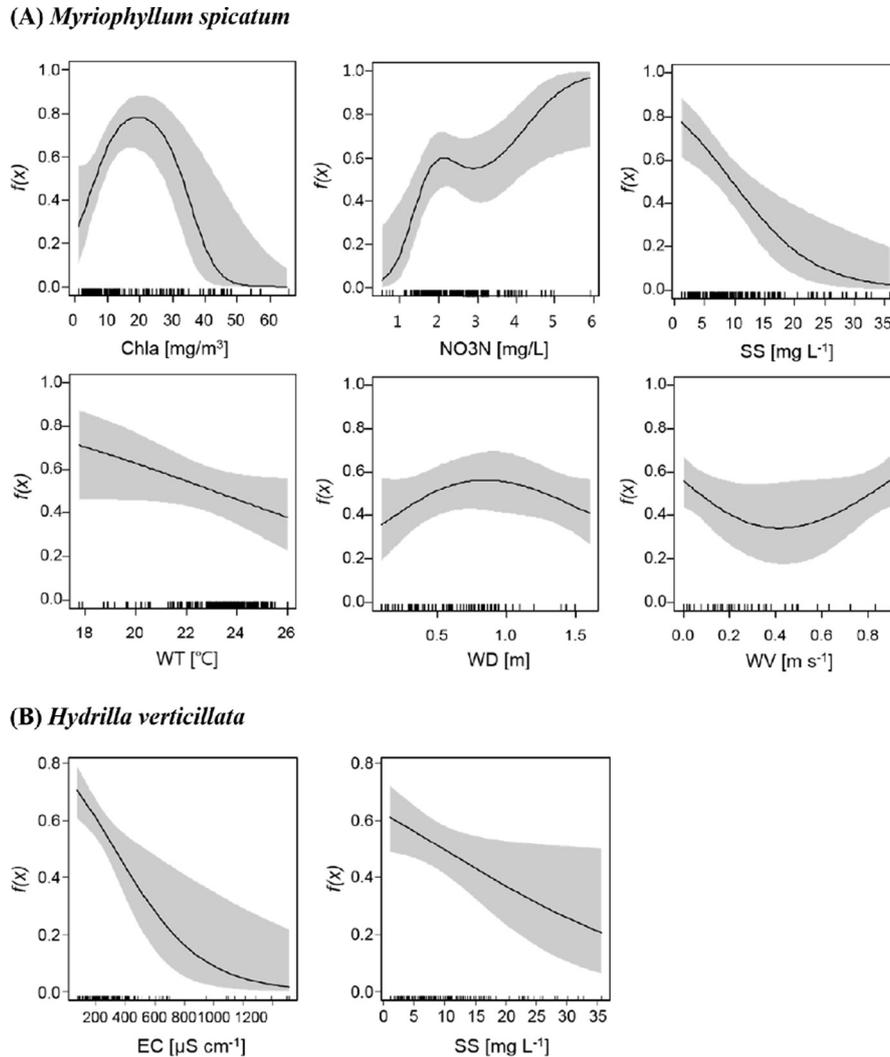
Wood, 2000), using the “mgcv” library for GAMs (Wood, 2006). The most parsimonious model for each species with the fewest variables was chosen using a stepwise selection procedure (Buisson *et al.*, 2008). To determine the best fit model, Akaike’s information criterion (AIC) was used as a goodness-of-fit statistic (Zuur and Pierce, 2004; Buisson *et al.*, 2008; Sanchez *et al.*, 2008). Models with a smaller AIC were able to explain the residual deviance better than those with a larger AIC. The best model accounted for the most variation in the data using the fewest predictors (Burnham and Anderson, 1998).

### 2.3 Validation of the predictive performance of models

To obtain an unbiased estimation of model performance, it is best to apply independent data that have not been used for model development (Pearce and Ferrier, 2000). However, if independent data are not available, a *k*-fold cross-validation may be used to assess model accuracy (Pearce and Ferrier, 2000; Zimmermann *et al.*, 2007). We applied *k*-fold cross-validation (with *k* = 10); the data were randomly split into two datasets: 90% (*k* – 1 subsamples) was used as a training dataset to build a model and the remaining 10% (one subsample) was used as a testing dataset for validation. This procedure was repeated ten times to calculate the probabilities of occurrence, which were transformed into binary records (presence/absence) using a threshold probability (Sing *et al.*, 2005; Araújo and Luoto, 2007). To

generate a confusion matrix, the Youden index was chosen as the threshold probability, which was the maximum difference between sensitivity (the probability of correct classification as positive) and specificity (the probability of correct classification as negative) (Jiménez-Valverde and Lobo, 2007; Freeman and Moisen, 2008), using the “SDMTools” library for model accuracy (VanDerWal *et al.*, 2014). Coordinates of observations and predictions based on the confusion matrix were projected to World Geodetic System 84 (WGS84) using QGIS (QGIS Development Team, 2016).

Model accuracy was assessed by two measures: Cohen’s kappa and the area under the receiver operating characteristic (ROC) curve (AUC) (Zimmermann *et al.*, 2007). Cohen’s kappa is the most common method for determining the accuracy of presence–absence predictions based on a selected threshold probability, regardless of variation in prevalence (Segurado and Araujo, 2004; Allouche *et al.*, 2006). Landis and Koch (1977) suggested the following interpretation of kappa values: excellent agreement, >0.75; good agreement, 0.40–0.75; and poor agreement, <0.40. Another method for assessing the accuracy of models uses the AUC value as a threshold-independent criterion (Fielding and Bell, 1997). To construct ROC curves, all possible thresholds were used to classify the scores into confusion matrices, and the sensitivity and specificity were estimated for each matrix (Allouche *et al.*, 2006). According to Swets (1988), AUC values were interpreted as follows: excellent, >0.90; good, 0.80–0.90; fair, 0.70–0.80; poor, 0.60–0.70; fail, 0.50–0.60. Moreover, *in*



**Fig. 2.** Response curves of (A) *Myriophyllum spicatum* and (B) *Hydrilla verticillata* for environmental gradients in GAMs. The vertical axes represent the probabilities of occurrence, and shaded bands show the 95% confidence interval. Rug plots on the x-axis show data points. Chla, chlorophyll *a*; EC, electrical conductivity; NO<sub>3</sub>N, nitrate nitrogen; SS, suspended solids; WT, water temperature; WD, water depth; WV, water velocity.

*situ* observations of the 41 sites were randomly performed to evaluate the accuracy of the model for field verification.

### 3 Results

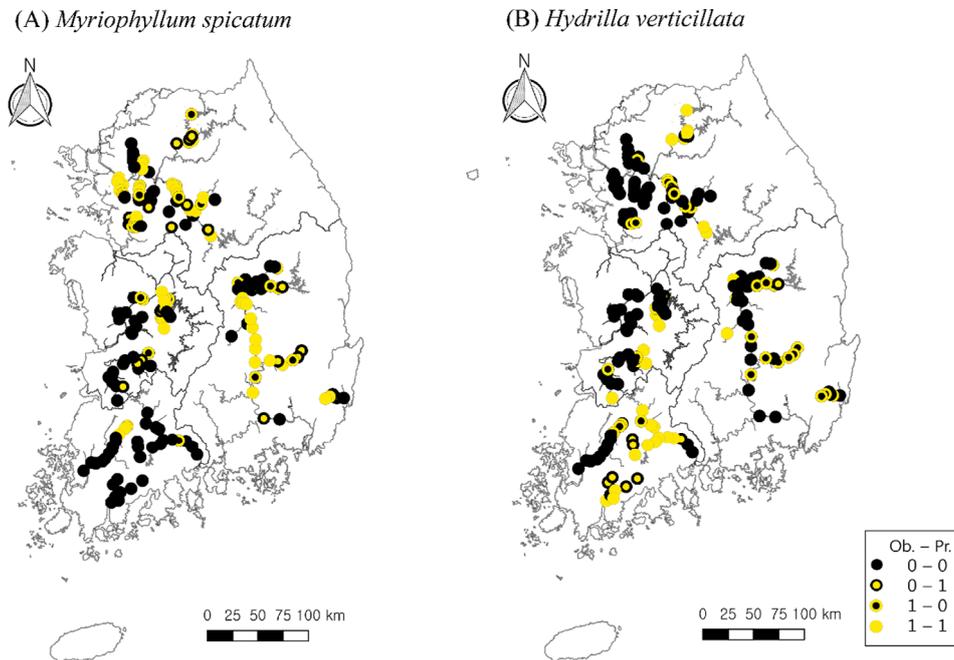
#### 3.1 GAM response curves

We defined the ranges of environmental factors with respect to species occurrences. The models explained 28.7% and 23.4% of the observed variation in selected variables for *M. spicatum* and *H. verticillata*, respectively (Tab. 2). For *M. spicatum*, we found that chlorophyll *a*, nitrate nitrogen, suspended solids, water temperature, water depth, and water velocity were significant variables in the GAM. Based on the response curves, the probability of *M. spicatum* presence increased as nitrate nitrogen increased, and decreased as water temperature and suspended solids increased (Fig. 2A). The response curves for chlorophyll *a* indicated a sharp increase in the predicted presence of *M. spicatum* from 0 to ~20 mg/m<sup>3</sup> and a decline at higher concentrations. The presence

probabilities of *M. spicatum* for water velocity decreased for values of up to ~0.4 m/s and then increased in relatively rapid flows. *Myriophyllum spicatum* was distributed in a wide range of water depths up to 1 m and decreased in deeper water. Electrical conductivity and suspended solids were important parameters determining the *H. verticillata* distribution (Tab. 2). *Hydrilla verticillata* presence was negatively related to electrical conductivity and suspended solid concentrations (Fig. 2B). We also observed differences in the predicted potential habitats between the two species (Fig. 3). *Myriophyllum spicatum* was widely distributed in the Han River and Nakdong River, and its predicted distribution was similar to its current distribution. *Hydrilla verticillata* was abundant everywhere, whereas its distribution was predicted to be decreased in Nakdong River and increased in Yeongsan River.

#### 3.2 Model validation and field verification

To assess model performances for each species, we compared predicted potential habitats with observed habitats



**Fig. 3.** Predicted and observed habitat suitability of (A) *Myriophyllum spicatum* and (B) *Hydrilla verticillata* based on GAMs. Observed occurrence (Ob.) is overlaid with the predicted occurrence (Pr.). The outer circle indicates the observed distribution, and the inner circle represents the predicted distribution. Yellow indicates presence (1) and black indicates absence (0). The same color for the outer and inner circles shows that observations and predictions coincide. A black outer circle with a yellow inner circle indicates a false positive, and a yellow outer with a black inner circle indicates a false negative.

using the confusion matrix (Tab. 3). We observed accuracy rates of 0.74 for *M. spicatum* and 0.75 for *H. verticillata*. For *M. spicatum*, we observed AUC=0.84 and kappa=0.46 when we applied a threshold of 0.536. For *H. verticillata*, we observed AUC=0.79 and kappa=0.39 when the threshold was 0.492. Based on the field verification, we observed accuracy rates of 76% for the two species (Tab. 4). In addition, we established AUC values of 0.75 for *M. spicatum* and 0.82 for *H. verticillata*, and kappa values of 0.44 for *M. spicatum* and 0.25 for *H. verticillata*.

## 4 Discussion

In this study, we identified factors that characterize the suitable habitats of *M. spicatum* and *H. verticillata* using GAMs, namely, chlorophyll *a*, electrical conductivity, nitrate nitrogen, suspended solids, water temperature, water depth, and water velocity. GAMs are very useful for describing the complex relationships between response variables and environmental factors (Cheng and Gallinat, 2004); however, our models explained low proportions of deviance (Tab. 2). This can be explained by the patchy distribution of submerged macrophytes, which cannot be fully explained by the selected variables (Lehmann, 1998). Nevertheless, based on the model evaluation procedures, the response curves for each variable enable us to infer general trends and we expect to adapt models to other locations around the world.

### 4.1 GAM results and environmental factors

The abundance and distribution of submerged macrophytes in river ecosystems are related to water quality conditions

(Nieder *et al.*, 2004), water depth, and water velocity (Sousa, 2011). We found that variables associated with water environmental factors were important determinants of the distributions of *M. spicatum* and *H. verticillata*, especially chlorophyll *a*, electrical conductivity, nitrate nitrogen, suspended solids, water temperature, water depth, and water velocity. Gradients of these variables also determine suitable habitats for submerged macrophytes in previous studies (Dodkins *et al.*, 2005; Lacoul and Freedman, 2006; Franklin *et al.*, 2008).

Our results demonstrated that the occurrence of *M. spicatum* increased for low concentrations, and decreased for high concentrations of chlorophyll *a*. In addition, the probability of *M. spicatum* presence decreased as water temperature increased. A higher temperature within optimal ranges usually promotes a higher chlorophyll *a* concentration and productivity as well as a greater abundance of submerged macrophytes (Barko *et al.*, 1986). However, competition for light between aquatic plants and phytoplankton may limit plant growth and even result in the disappearance of taxa (Rybicki and Landwehr, 2007; Bornette and Puijalón, 2011). Jones *et al.* (1983) also found that phytoplankton-created low water clarity inhibits *M. spicatum* growth. Furthermore, the adaptability of submerged macrophytes to low temperatures may play a role in interspecific competition because the optimal water temperature for submerged macrophytes is 28–32 °C (Barko *et al.*, 1986).

The occurrence of *H. verticillata* was high when electrical conductivity was low in oligotrophic water (Fig. 2B). This agrees with the findings of Sousa *et al.* (2009) and Cook and Lüönd (1982) who found *H. verticillata* develops well in oligotrophic waters. Plant growth usually increases as the

**Table 3.** Comparison of predicted and observed distributions of *Myriophyllum spicatum* and *Hydrilla verticillata* during model building. Predicted values were obtained from the fitted probability of presence using the Youden index to apply a threshold probability: 0.536 for *M. spicatum* and 0.492 for *H. verticillata*.

Species	Contents	Predicted absence	Predicted presence	Total
<i>M. spicatum</i>	Observed absence	91	23	114
	Observed presence	29	54	83
	Total	120	77	197
	Correct prediction	$(91 + 54)/197 = 0.74$		
	Error of commission	$23/114 = 0.20$		
	Error of omission	$29/83 = 0.35$		
	AUC	0.84		
	Kappa	0.46		
<i>H. verticillata</i>	Observed absence	115	13	128
	Observed presence	37	32	69
	Total	152	45	197
	Correct prediction	$(115 + 32)/197 = 0.75$		
	Error of commission	$13/128 = 0.10$		
	Error of omission	$37/69 = 0.54$		
	AUC	0.79		
	Kappa	0.39		

concentration of nutrients in water and sediment increases (Van *et al.*, 1999; Yu *et al.*, 2010; Sousa, 2011). Moreover, Kennedy *et al.* (2009) found that *H. verticillata* can thrive not only in eutrophic waters, but also in oligotrophic waters. According to Sousa (2011), however, eutrophic conditions may have negative effects on *H. verticillata* growth via the proliferation of plankton, which compete with submerged macrophytes for light and nutrients. Electrical conductivity, as a measure of the chemicals summary variable (Heegaard *et al.*, 2001), may affect macrophyte composition and be unfavorable for submerged macrophytes that are sensitive to eutrophication (Thomaz *et al.*, 2003; Lauridsen *et al.*, 2015). The occurrence of *M. spicatum* was high when nitrate nitrogen was high in water (Fig. 2A). Nitrogen is a key element for aquatic plants, which use nitrate as a nitrogen source (Bornette and Puijalon, 2011) and *M. spicatum* was related to with higher nitrate availability (Ali and Soltan, 2006).

We detected abundant *M. spicatum* in water at depths of 0.7–1.0 m, and a decreased abundance in deeper water (Fig. 2A). This result concurs with a previous study; Nichols and Shaw (1986) reported that *M. spicatum* is typically found in water ranging from 1 to 4 m deep. Angradi *et al.* (2013) observed that the optimal depth for submerged macrophytes is ~1.2 m and <1 m in turbid conditions. However, an increase in water depth causes a light deficiency for submerged macrophytes (Bornette and Puijalon, 2011), thereby decreasing the rate of photosynthesis. Shallow water bodies allow more light penetration, provided that the water is not turbid (Narumalani *et al.*, 1997). Lower water clarity owing to sediments, turbidity, and nutrients can reduce the water depth and spatial distribution for growth and survival of submerged macrophytes (Dar *et al.*, 2014; Patrick *et al.*, 2014). Consequently, as depth increases or water clarity decreases, light availability for photosynthesis may diminish (Lacoul and Freedman, 2006).

Based on the velocity response curve, we observed a decrease in *M. spicatum* abundance for values of up to ~0.4 m/s

and an increase for higher values (Fig. 2A). On the other hand, *M. spicatum* has been found in the stream at water velocities of 2 m/s (Nichols and Shaw, 1986). A number of water velocity readings in this study were at 0 m/s, with an average of 0.1 m/s and a median of 0 m/s. The regulated rivers and streams in study sites were close to lentic conditions; accordingly, we could not precisely evaluate the velocity response of *M. spicatum*. However, in general, the biomass and richness of submerged macrophytes are higher at 0.3–0.4 m/s, and lower at higher velocities because of failure to establish and colonize (Lacoul and Freedman, 2006). Moderate flow can encourage submerged macrophyte growth by continuously washing photosynthetic tissues covered with epiphytic algae (Strand and Weisner, 1996; Lehmann, 1998).

## 4.2 Model validation and field verification

We did not develop GIS-based predictive maps to identify potential distribution for submerged macrophytes because it was difficult to construct a bathymetric map of all rivers and streams in South Korea. We present maps showing areas of agreement between observations and predictions. Although we did not generate spatial maps interpolated with predicted probabilities, we were able to detect areas with abundant submerged macrophytes in the five rivers and characterize the suitable habitats of them.

Correct and incorrect predictions in a confusion matrix indicate the strength of predictions (Peters *et al.*, 2007). The misclassified sites in the predicted distribution according to GAMs were related to channel structure, rather than water quality. Most of these sites were confluence points, *i.e.*, sites at which two channels met, each carrying independent influxes and sediment discharge (Benda *et al.*, 2004). To examine false positive errors (commission error; observation=0 and prediction=1), we considered the characteristics of survey sites at confluence points that connected relatively larger

**Table 4.** Comparison of predicted and observed distributions of *Myriophyllum spicatum* and *Hydrilla verticillata* at the model confirmation stage. Predicted values were obtained from the fitted probability of presence using the Youden index to apply a threshold probability: 0.536 for *M. spicatum* and 0.492 for *H. verticillata*.

Species	Contents	Predicted absence	Predicted presence	Total
<i>M. spicatum</i>	Observed absence	23	6	29
	Observed presence	4	8	12
	Total	27	14	41
	Correct prediction	$(23 + 8)/41 = 0.76$		
	Error of commission	$6/29 = 0.21$		
	Error of omission	$4/12 = 0.33$		
	AUC	0.75		
	Kappa	0.44		
<i>H. verticillata</i>	Observed absence	28	2	30
	Observed presence	8	3	11
	Total	36	5	41
	Correct prediction	$(28 + 3)/41 = 0.76$		
	Error of commission	$2/30 = 0.07$		
	Error of omission	$8/11 = 0.73$		
	AUC	0.82		
	Kappa	0.25		

tributaries to main water bodies. Confluences have been described as biodiversity hot spots with physical heterogeneity and habitat complexity owing to diverse physical, chemical, and biological attributes resulting from tributary streams (Kiffney *et al.*, 2006; Rice *et al.*, 2006). At these sites, we expected to observe submerged macrophytes; however, their distributions were discontinuous and bed sediment size and flow properties were unstable (Rice *et al.*, 2006). Slow velocity conditions and the maintenance of stable substrates are necessary for macrophytes colonization (Riis and Biggs, 2003). False positive results imply that submerged macrophytes did not have an opportunity to establish to a suitable habitat (Buchan and Padilla, 2000).

We observed false negative errors (omission error; observation=1 and prediction=0) at study sites that were typically downstream of confluence points connecting relatively small tributaries and irrigation ditches with the potential for submerged macrophytes dispersal. *Myriophyllum spicatum* and *H. verticillata* are found in rivers, lakes, irrigation ditches, and other waterways (Netherland, 1997; Eiswerth *et al.*, 2000). Their primary dispersal strategy is vegetative reproduction by fragmentation, moving through small waterways and then establishing at channel junctions. Vegetative reproduction by stem fragmentation is an efficient mechanism for dispersal, colonization, and overwintering (Sousa, 2011; Xie *et al.*, 2013). The conditions downstream of small tributaries are sufficiently stable for submerged macrophytes inhabitation. Submerged macrophytes habitats are regulated by a variety of factors, varying within not only whole streams, but also smaller stream reaches (Riis *et al.*, 2001).

We observed high accuracy rates for each species model (0.74 for *M. spicatum* and 0.75 for *H. verticillata*), but low proportions of variation explained by the models. We observed kappa values (which were dependent on a threshold) of 0.46 for *M. spicatum* and 0.39 for *H. verticillata*, indicating a fair

model fit. In addition, we observed AUC values (independent of threshold values) of 0.84 for *M. spicatum* and 0.79 for *H. verticillata*, indicating satisfactory predictive ability. According to the AUC and kappa values, model performance was good for both species. Field verification to validate favorable potential habitats for *M. spicatum* and *H. verticillata* confirmed model performance, supporting their good prediction abilities, based on accuracy rates, AUC, and kappa values, except for the kappa value (0.25) of *H. verticillata*.

Although GAMs do not provide superior predictive performance compared with other models such as multivariate adaptive regression splines and boosted regression trees (Austin, 2007), they are flexible enough to model relationships between occurrences of submerged macrophytes and environmental factors (Murase *et al.*, 2009). Habitats for submerged macrophytes are characterized by a complex set of physical, chemical, and biological parameters. In this study, the modeling of potential habitats for submerged macrophytes was in good agreement, despite only considering water chemicals, water depth, and water velocity. However, a reasonable possibility of prediction errors is the reason why the realized niches of submerged macrophytes were not completely explained by the variables that we selected and added to GAMs. We measured water depth and water velocity once in normal conditions, and we used water chemical data that were averaged over four years. Even though a lack of long-term monitoring data for water depth and velocity at the study sites is a source of uncertainty, we were able to overcome uncertainties by surveying diverse environmental conditions, varying from tributaries to rivers.

Buchan and Padilla (2000) found that the most important factors affecting the presence-absence of submerged macrophytes were associated with water quality rather than with dispersal potential. Nonetheless, to improve the predictive accuracy, it is also necessary to consider physical factors, such

as flow regime, channel connectivity, channel slope, channel bed, shoreline conditions, and land cover type in the basin (Buchan and Padilla, 2000; Patrick *et al.*, 2014). Biological factors, including competition, herbivory, and disease, are also important habitat determinants (Lacoul and Freedman, 2006). No habitat suitability model is a complete representation of reality and these models should be validated for applications using real-world data by predictive performance evaluations focusing on the reduction of omission errors (Liu *et al.*, 2009; Gastón and García-Viñas, 2013). The distribution of *M. spicatum* and *H. verticillata* with high invasiveness could not be compared before and after water regulation in South Korea; however, we are able to predict the distribution using water quality data and adapt the models throughout the world.

### 4.3 Perspectives

We inferred the habitat characteristics of *M. spicatum* and *H. verticillata* using GAMs based on field survey data at the catchment scale. We observed that water chemicals, *e.g.*, chlorophyll *a*, suspended solids, nitrate nitrogen, water temperature, and electrical conductivity are important factors determining the occurrences of submerged macrophytes. Moreover, we validated the developed models by applying independent field data. Our research has practical implications for the prevention or delay of the aggressive spread of *M. spicatum* and *H. verticillata* by providing a basis for river management strategies, such as information about water chemicals to improve water quality in priority areas (Barko *et al.*, 1986). These results are also helpful to sustain aquatic ecosystem functions and biodiversity in regulated hydrological conditions by identifying priority areas for monitoring and management.

### Supplementary Material

Supplementary Tables.

The Supplementary Material is available at <https://www.kmae-journal.org/10.1051/kmae/2017044/olm>.

### References

- Ahmadi-Nedushan B, St-Hilaire A, Bérubé M, Robichaud É, Thiémonge N, Bobée B. 2006. A review of statistical methods for the evaluation of aquatic habitat suitability for instream flow assessment. *River Res Appl* 22: 503–523.
- Ali MM, Soltan MA. 2006. Expansion of *Myriophyllum spicatum* (Eurasian water milfoil) into Lake Nasser, Egypt: invasive capacity and habitat stability. *Aquat Bot* 84: 239–244.
- Allouche O, Tsoar A, Kadmon R. 2006. Assessing the accuracy of species distribution models: prevalence, kappa and the true skill statistic (TSS). *J Appl Ecol* 43: 1223–1232.
- Angradi TR, Pearson MS, Bolgrien DW, Bellinger BJ, Starry MA, Reschke C. 2013. Predicting submerged aquatic vegetation cover and occurrence in a Lake Superior estuary. *J Great Lakes Res* 39: 536–546.
- Araújo MB, Luoto M. 2007. The importance of biotic interactions for modelling species distributions under climate change. *Glob Ecol Biogeogr* 16: 743–753.
- Austin MP. 2002. Spatial prediction of species distribution: an interface between ecological theory and statistical modelling. *Ecol Model* 157: 101–118.
- Austin MP. 2007. Species distribution models and ecological theory: a critical assessment and some possible new approaches. *Ecol Model* 200: 1–19.
- Austin MP, Meyers JA. 1996. Current approaches to modelling the environmental niche of eucalypts: implication for management of forest biodiversity. *For Ecol Manage* 85: 95–106.
- Barko J, Adams M, Clesceri N. 1986. Environmental factors and their consideration in the management of submerged aquatic vegetation: a review. *J Aquat Plant Manage* 24: 1–10.
- Beck KG, Zimmerman K, Schardt JD, *et al.* 2008. Invasive species defined in a policy context: recommendations from the Federal Invasive Species Advisory Committee. *Invasive Plant Sci Manage* 1: 414–421.
- Benda L, Andras K, Miller D, Bigelow P. 2004. Confluence effects in rivers: interactions of basin scale, network geometry, and disturbance regimes. *Water Resour Res* 40: W05402.
- Borchers DL, Buckland ST, Priede IG, Ahmadi S. 1997. Improving the precision of the daily egg production method using generalized additive models. *Can J Fish Aquat Sci* 54: 2727–2742.
- Bornette G, Puijalon S. 2011. Response of aquatic plants to abiotic factors: a review. *Aquat Sci* 73: 1–14.
- Bučas M, Bergström U, Downie A-L, *et al.* 2013. Empirical modelling of benthic species distribution, abundance, and diversity in the Baltic Sea: evaluating the scope for predictive mapping using different modelling approaches. *ICES J Mar Sci* 70: 1233–1243.
- Buchan LAJ, Padilla DK. 2000. Predicting the likelihood of Eurasian watermilfoil presence in lakes, a macrophyte monitoring tool. *Ecol Appl* 10: 1442–1455.
- Buisson L, Blanc L, Grenouillet G. 2008. Modelling stream fish species distribution in a river network: the relative effects of temperature versus physical factors. *Ecol Freshwater Fish* 17: 244–257.
- Bunn SE, Arthington AH. 2002. Basic principles and ecological consequences of altered flow regimes for aquatic biodiversity. *Environ Manage* 30: 492–507.
- Burnham K, Anderson D. 1998. Model selection and inference: a practical information-theoretic approach. New York: Springer.
- Camporeale C, Ridolfi L. 2006. Riparian vegetation distribution induced by river flow variability: a stochastic approach. *Water Resour Res* 42: W10415.
- Canfield D, Langeland K, Linda S, Haller W. 1985. Relations between water transparency and maximum depth of macrophyte colonization in lakes. *J Aquat Plant Manage* 23: 25–28.
- Chambers PA, DeWreede RE, Irlandi EA, Vandermeulen H. 1999. Management issues in aquatic macrophyte ecology: a Canadian perspective. *Can J Bot* 77: 471–487.
- Chappuis E, Gacia E, Ballesteros E. 2014. Environmental factors explaining the distribution and diversity of vascular aquatic macrophytes in a highly heterogeneous Mediterranean region. *Aquat Bot* 113: 72–82.
- Cheng YW, Gallinat MP. 2004. Statistical analysis of the relationship among environmental variables, inter-annual variability and smolt trap efficiency of salmonids in the Tucannon River. *Fish Res* 70: 229–238.
- Clayton J, Edwards T. 2006. Aquatic plants as environmental indicators of ecological condition in New Zealand lakes. *Hydrobiologia* 570: 147–151.
- Cook CDK, Lüönd R. 1982. A revision of the genus *Hydrilla* (Hydrocharitaceae). *Aquat Bot* 13: 485–504.

- Dar NA, Pandit AK, Ganai BA. 2014. Factors affecting the distribution patterns of aquatic macrophytes. *Limnol Rev* 14: 75–81.
- Dawson FH, Raven PJ, Gravelle MJ. 1999. Distribution of the morphological groups of aquatic plants for rivers in the U.K. In Caffrey J, Barrett PRF, Ferreira MT, Moreira IS, Murphy KJ, Wade PM, eds. *Biology, Ecology and Management of Aquatic Plants: Proceedings of the 10th International Symposium on Aquatic Weeds*, European Weed Research Society. Dordrecht: Springer Netherlands, pp. 123–130.
- Dennison WC, Orth RJ, Moore KA, *et al.* 1993. Assessing water quality with submersed aquatic vegetation. *Bioscience* 43: 86–94.
- Dodkins IAN, Rippey B, Hale P. 2005. An application of canonical correspondence analysis for developing ecological quality assessment metrics for river macrophytes. *Freshw Biol* 50: 891–904.
- Downie AL, von Numers M, Boström C. 2013. Influence of model selection on the predicted distribution of the seagrass *Zostera marina*. *Estuar Coast Shelf Sci* 121: 8–19.
- Drexler M, Ainsworth CH. 2013. Generalized additive models used to predict species abundance in the Gulf of Mexico: an ecosystem modeling tool. *PLoS ONE* 8: e 64458.
- Dynesius M, Nilsson C. 1994. Fragmentation and flow regulation of river systems in the Northern Third of the World. *Science* 266: 753–762.
- Eiswerth ME, Donaldson SG, Johnson WS. 2000. Potential environmental impacts and economic damages of Eurasian watermilfoil (*Myriophyllum spicatum*) in Western Nevada and Northeastern California. *Weed Technol* 14: 511–518.
- Eliith J, Graham CH, Anderson RP, *et al.* 2006. Novel methods improve prediction of species' distributions from occurrence data. *Ecography* 29: 129–151.
- Fielding AH, Bell JF. 1997. A review of methods for the assessment of prediction errors in conservation presence/absence models. *Environ Conserv* 24: 38–49.
- Franklin P, Dunbar M, Whitehead P. 2008. Flow controls on lowland river macrophytes: a review. *Sci Total Environ* 400: 369–378.
- Freeman EA, Moisen GG. 2008. A comparison of the performance of threshold criteria for binary classification in terms of predicted prevalence and kappa. *Ecol Model* 217: 48–58.
- Gallego I, Pérez-Martínez C, Sánchez-Castillo PM, Fuentes-Rodríguez F, Juan M, Casas JJ. 2015. Physical, chemical, and management-related drivers of submerged macrophyte occurrence in Mediterranean farm ponds. *Hydrobiologia* 762: 209–222.
- Gassmann A, Cock MJW, Shaw R, Evans HC. 2006. The potential for biological control of invasive alien aquatic weeds in Europe: a review. *Hydrobiologia* 570: 217–222.
- Gastón A, García-Viñas JI. 2013. Evaluating the predictive performance of stacked species distribution models applied to plant species selection in ecological restoration. *Ecol Model* 263: 103–108.
- Guisan A, Edwards Jr TC, Hastie T. 2002. Generalized linear and generalized additive models in studies of species distributions: setting the scene. *Ecol Model* 157: 89–100.
- Heegaard E, Birks HH, Gibson CE, Smith SJ, Wolfe-Murphy S. 2001. Species-environment relationships of aquatic macrophytes in Northern Ireland. *Aquat Bot* 70: 175–223.
- Järvelä J. 2005. Effect of submerged flexible vegetation on flow structure and resistance. *J Hydrol* 307: 233–241.
- Jiménez-Valverde A, Lobo JM. 2007. Threshold criteria for conversion of probability of species presence to either-or presence-absence. *Acta Oecol* 31: 361–369.
- Jones RC, Walti K, Adams MS. 1983. Phytoplankton as a factor in the decline of the submersed macrophyte *Myriophyllum spicatum* L. in Lake Wingra, Wisconsin, USA. *Hydrobiologia* 107: 213–219.
- Jun KS, Kim JS. 2011. The four major rivers restoration project: impacts on river flows. *KSCE J Civ Eng* 15: 217–224.
- Kennedy TL, Horth LA, Carr DE. 2009. The effects of nitrate loading on the invasive macrophyte *Hydrilla verticillata* and two common, native macrophytes in Florida. *Aquat Bot* 91: 253–256.
- Kenneth AL. 1996. *Hydrilla verticillata* (L.F.) Royle (Hydrocharitaceae), “The Perfect Aquatic Weed”. *Castanea* 61: 293–304.
- Kiffney PM, Greene CM, Hall J, Davies J. 2006. Tributary streams create spatial discontinuities in habitat, biological productivity, and diversity in mainstem rivers. *Can J Fish Aquat Sci* 63: 2518–2530.
- Klippel S, Amaral S, Vinhas L. 2016. Development and evaluation of species distribution models for five endangered elasmobranchs in southwestern Atlantic. *Hydrobiologia* 779: 11–33.
- Kuhn M, Johnson K. 2013. *Applied predictive modeling*. New York: Springer.
- Lacoul P, Freedman B. 2006. Environmental influences on aquatic plants in freshwater ecosystems. *Environ Rev* 14: 89–136.
- Lah T, Park Y, Cho YJ. 2015. The four major rivers restoration project of South Korea: an assessment of its process, program, and political dimensions. *J Environ Dev* 24: 375–394.
- Landis JR, Koch GG. 1977. The measurement of observer agreement for categorical data. *Biometrics* 33 159–174.
- Lathrop RG, Styles RM, Seitzinger SP, Bognar JA. 2001. Use of GIS mapping and modeling approaches to examine the spatial distribution of seagrasses in Barnegat Bay, New Jersey. *Estuaries* 24: 904–916.
- Lauridsen TL, Jeppesen E, Declerck SAJ, *et al.* 2015. The importance of environmental variables for submerged macrophyte community assemblage and coverage in shallow lakes: differences between northern and southern Europe. *Hydrobiologia* 744: 49–61.
- Lee JW, Bae SI, Lee DR, Seo DI. 2014. Transportation modeling of conservative pollutant in a river with weirs – the Nakdong River case. *J Korean Soc Environ Eng* 36: 821–827.
- Lehmann A. 1998. GIS modeling of submerged macrophyte distribution using Generalized Additive Models. *Plant Ecol* 139: 113–124.
- Li X, Wang Y. 2013. Applying various algorithms for species distribution modelling. *Integr Zool* 8: 124–135.
- Liu C, White M, Newell G. 2009. Measuring the accuracy of species distribution models: a review. In: *Proceedings 18th World IMACS/ MODSIM Congress, Cairns, Australia*, pp. 4241–4247.
- Murase H, Nagashima H, Yonezaki S, Matsukura R, Kitakado T. 2009. Application of a generalized additive model (GAM) to reveal relationships between environmental factors and distributions of pelagic fish and krill: a case study in Sendai Bay, Japan. *ICES J Mar Sci* 66: 1417–1424.
- Narumalani S, Jensen JR, Burkhalter S, Althausen JD, Mackey Jr HE. 1997. Aquatic macrophyte modeling using GIS and logistic multiple regression. *Photogramm Eng Remote Sens* 63: 41–49.
- Nepf H, Ghisalberti M, White B, Murphy E. 2007. Retention time and dispersion associated with submerged aquatic canopies. *Water Resour Res* 43: W04422.
- Netherland MD. 1997. Turion ecology of hydrilla. *J Aquat Plant Manage* 35: 1–10.
- Nichols SA, Shaw BH. 1986. Ecological life histories of the three aquatic nuisance plants, *Myriophyllum spicatum*, *Potamogeton crispus* and *Elodea canadensis*. *Hydrobiologia* 131: 3–21.
- Nieder WC, Barnaba E, Findlay SEG, Hoskins S, Holochuck N, Blair EA. 2004. Distribution and abundance of submerged aquatic vegetation and *Trapa natans* in the Hudson River Estuary. *J Coast Res* 150–161.
- Normile D. 2010. Restoration or devastation? *Science* 327: 1568–1570.
- O'Hare JM, O'Hare MT, Gurnell AM, Dunbar MJ, Scarlett PM, Laizé C. 2011. Physical constraints on the distribution of macrophytes

- linked with flow and sediment dynamics in British rivers. *River Res Appl* 27: 671–683.
- Park HK, Cho KH, Won DH, Lee J, Kong DS, Jung DI. 2013. Ecosystem responses to climate change in a large on-river reservoir, Lake Paldang, Korea. *Clim Change* 120: 477–489.
- Patrick CJ, Weller DE, Li X, Ryder M. 2014. Effects of shoreline alteration and other stressors on submerged aquatic vegetation in subestuaries of Chesapeake Bay and the Mid-Atlantic Coastal Bays. *Estuar Coast* 37: 1516–1531.
- Pearce J, Ferrier S. 2000. Evaluating the predictive performance of habitat models developed using logistic regression. *Ecol Model* 133: 225–245.
- Peters J, Baets BD, Verhoest NEC, Samson R, Degroeve S, Becker PD, Huybrechts W. 2007. Random forests as a tool for ecohydrological distribution modelling. *Ecol Model* 207: 304–318.
- QGIS Development Team. 2016. QGIS Geographic Information System. *Open Source Geospatial Foundation Project*. <http://www.qgis.org/>.
- R Development Core Team. 2016. R: A language and environment for statistical computing. *Vienna: R Foundation for Statistical Computing*. <http://www.r-project.org/>.
- Rice SP, Ferguson RI, Hoey TB. 2006. Tributary control of physical heterogeneity and biological diversity at river confluences. *Can J Fish Aquat Sci* 63: 2553–2566.
- Riis T, Biggs B. 2003. Hydrologic and hydraulic control of macrophyte establishment and performance in streams. *Limnol Oceanogr* 48: 1488–1497.
- Riis T, Sand-Jensen K, Larsen SE. 2001. Plant distribution and abundance in relation to physical conditions and location within Danish stream systems. *Hydrobiologia* 448: 217–228.
- Rybicki NB, Landwehr JM. 2007. Long-term changes in abundance and diversity of macrophyte and waterfowl populations in an estuary with exotic macrophytes and improving water quality. *Limnol Oceanogr* 52: 1195–1207.
- Sanchez P, Demestre M, Recasens L, Maynou F, Martin P. 2008. Combining GIS and GAMs to identify potential habitats of squid *Loligo vulgaris* in the Northwestern Mediterranean. *Hydrobiologia* 612: 91–98.
- Segurado P, Araujo MB. 2004. An evaluation of methods for modelling species distributions. *J Biogeogr* 31: 1555–1568.
- Shin JH, Chung JY. 2011. The four major rivers restoration project in South Korea. *Proc Inst Civ Eng Civ Eng* 164: 19–26.
- Shmueli G. 2010. To explain or to predict? *Stat Sci* 289–310.
- Sing T, Sander O, Beerwinkel N, Lengauer T. 2005. ROCR: visualizing classifier performance in R. *Bioinformatics* 21: 3940–3941.
- Solanki HU, Bhatpuria D, Chauhan P. 2016. Applications of generalized additive model (GAM) to satellite-derived variables and fishery data for prediction of fishery resources distributions in the Arabian Sea. *Geocarto Int* 32: 30–43.
- Sousa W. 2011. *Hydrilla verticillata* (Hydrocharitaceae), a recent invader threatening Brazil's freshwater environments: a review of the extent of the problem. *Hydrobiologia* 669: 1–20.
- Sousa WTZ, Thomaz SM, Murphy KJ, Silveira MJ, Mormul RP. 2009. Environmental predictors of the occurrence of exotic *Hydrilla verticillata* (L.f.) Royle and native *Egeria najas* Planch. in a sub-tropical river floodplain: the Upper River Paraná, Brazil. *Hydrobiologia* 632: 65–78.
- Strand JA, Weisner SEB. 1996. Wave exposure related growth of epiphyton: implications for the distribution of submerged macrophytes in eutrophic lakes. *Hydrobiologia* 325: 113–119.
- Swets J. 1988. Measuring the accuracy of diagnostic systems. *Science* 240: 1285–1293.
- Thomaz SM, Souza DC, Bini LM. 2003. Species richness and beta diversity of aquatic macrophytes in a large subtropical reservoir (Itaipu Reservoir, Brazil): the influence of limnology and morphometry. *Hydrobiologia* 505: 119–128.
- Thuiller W, Richardson DM, Pyšek P, Midgley GF, Hughes GO, Rouget M. 2005. Niche-based modelling as a tool for predicting the risk of alien plant invasions at a global scale. *Glob Change Biol* 11: 2234–2250.
- Tian K, Liu G, Xiao D, Sun J, Lu M, Huang Y, Lin P. 2015. Ecological effects of dam impoundment on closed and half-closed wetlands in China. *Wetlands* 35: 889–898.
- Van TK, Wheeler GS, Center TD. 1999. Competition between *Hydrilla verticillata* and *Vallisneria spiralis* americana as influenced by soil fertility. *Aquat Bot* 62: 225–233.
- VanDerWal J, Falconi L, Januchowski S, Shoo L, Storlie C. 2014. SDMTools: Species Distribution Modelling tools: tools for processing data associated with species distribution modelling exercises. R package version, 1.1-221.
- Wedding L, Yoklavich MM. 2015. Habitat-based predictive mapping of rockfish density and biomass off the central California coast. *Mar Ecol Prog Ser* 540: 235–250.
- Woo H. 2010. Trends in ecological river engineering in Korea. *J Hydroenviron Res* 4: 269–278.
- Wood SN. 2000. Modelling and smoothing parameter estimation with multiple quadratic penalties. *J R Stat Soc Ser B (Stat Method)* 62: 413–428.
- Wood SN. 2006. *Generalized additive models: an introduction with R*. Boca Raton: CRC Press.
- Xie D, Yu D, You WH, Wang LG. 2013. Morphological and physiological responses to sediment nutrients in the submerged macrophyte *Myriophyllum spicatum*. *Wetlands* 33: 1095–1102.
- Yee TW, Mitchell ND. 1991. Generalized additive models in plant ecology. *J Veg Sci* 2: 587–602.
- Yu H, Ye C, Song X, Liu J. 2010. Comparative analysis of growth and physio-biochemical responses of *Hydrilla verticillata* to different sediments in freshwater microcosms. *Ecol Eng* 36: 1285–1289.
- Zhao J, Cao J, Tian S, Chen Y, Zhang S, Wang Z, Zhou X. 2014. A comparison between two GAM models in quantifying relationships of environmental variables with fish richness and diversity indices. *Aquat Ecol* 48: 297–312.
- Zhou N, Hu W, Deng J, Zhu J, Xu W, Liu X. 2016. The effects of water depth on the growth and reproduction of *Potamogeton crispus* in an in situ experiment. *J Plant Ecol* 10: 546–558.
- Zimmermann NE, Edwards TC, Moisen GG, Frescino TS, Blackard JA. 2007. Remote sensing-based predictors improve distribution models of rare, early successional and broadleaf tree species in Utah. *J Appl Ecol* 44: 1057–1067.
- Zuur AF, Pierce GJ. 2004. Common trends in northeast Atlantic squid time series. *J Sea Res* 52: 57–72.

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